

Kinetics of nitrobenzene biotransformation

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Abstract

Nitrobenzene and substituted aromatics like nitro-chlorobenzene, aniline, and nitro-toluene containing chlorine, amino and methyl groups are recalcitrant in nature. These molecules are both anthropogenic and xenobiotic. Nitro benzene and nitro aromatic compounds are environmental pollutants discharged through wastewaters from nitro aromatic manufacturing plants. Nitrobenzene and other aromatics are toxic to several forms of aquatic life. However, biological transformation of nitrobenzene to non-toxic entities exists in specialized microbes, which have enzymes of aromatic catabolic pathways. Transformation of nitrobenzene is also inhibited by the presence of other toxic materials such as cyanides and sulphides that are present in industrial waste when nitrobenzene is the dominant carbonaceous material. Kinetics of the biotransformation of nitrobenzene using pure cultures isolated in the laboratory, mixture of consortium of the pooled cultures and enriched activated sludge biomass in pure substrates, mixed substrates and actual nitrobenzene plant waste has been estimated under varying input concentrations. Experimental results from these studies have been subjected to analysis by mathematical models using Monod's and Haldane's equations to test their validity in interpreting the data on inhibitory substances under stable as well as unstable state operations of wastewater treatment plants. Comparative evaluation of the kinetic parameters reveals that Monod's model can be employed for the estimation of kinetic constant μ only while Haldane's model has to be used for the calculation of μ_{\max} and K_i .

Keywords: *kinetics, nitrobenzene, nitrochlorobenzene, and biotransformation.*



1 Introduction

Nitrobenzene is a substituted aromatic molecule containing nitro group. The origin of nitrobenzene to the atmosphere is due to anthropogenic activity. Anthropogenic sources are industrial wastes derived from chemical manufacturing processes such as nitro aromatic chemicals production units. Mostly the industrial wastes are originated from the wash waters of final product purification. The summary of microbial transformations of some substituted aromatic compounds together with isolated microorganisms is given in a research paper (Gibson [1]). Anaerobic degradation studies on benzene nucleus were carried out (Taylor *et al.* [2]) by using facultative and anaerobic microorganisms

Activated sludge process (ASP) is one of the most widely accepted biological systems for the treatment of nitro aromatic compounds. The biodegradation of aromatic pollutants under aerobic conditions was studied (Arcangeli and Arvin [3]). Existing theory suggests that operational difficulties associated with inhibitory compounds present in the waste is due to process dynamics. The significance of inhibitory mechanisms, associated with operational problems in the waste treatment, is to avoid metastable and unstable regimes. Bacteria capable of utilizing nitro aromatic compounds are found in soil and water environment. The wastewater from nitro aromatic production units predominantly contains nitrobenzene, nitrochlorobenzene, nitro toluene, and nitro cresols. Microorganisms attack these aromatics ring through a degradative pathway. The kinetics of microbial degradation of toluene were studied (Jirgensen *et al.* [4]). Aromatic compounds are rich in carbon content and once the rings are cleaved by enzymes the products (organic acids) enter the energy cycle. During the studies of biodegradation of nitro aromatic compounds it was observed (Hu and Sheih [5]) that nitro groups substituted by hydroxyl groups, with the elimination of the group as nitrite ion. Kinetics of the enzymatic reactions is hence significant in understanding the biotransformation of these molecules. Techniques for obtaining bio kinetic parameter values were studied (Grady [6]).

Table 1: Physico chemical characteristics of nitrobenzene wastewater.

S.No	Parameter	Value
1	PH	4.5-5.6
2	Color	Reddish brown
3	C.O.D.	1800-2200
4	Total solids	100-120
5	B.O.D.	650-800
6	Nitrobenzene	80-150
7	Suspended solids	50-80
8	Ammonical nitrogen	120-200
9	Oil and Grease	10-25

All values expressed in mg/L except pH and colour.



2 Theory

Evaluation of bio kinetic constants is significant for understanding the capacities of microorganisms for the operation of biological reactors. Kinetic studies on bacterial growth have given rise to considerable amount of literature. Particularly with regard to application of Monod's equation in respect of batch results.

$$\mu = \mu_{\max} \left(\frac{S}{K_s + S} \right) \quad (1)$$

In calculation of kinetic constants, μ , μ_{\max} , and K are related to reactor functions. These constants have been derived from the original Michaelis-Menton's enzyme equation:

$$v_0 = V_{\max} \left(\frac{S}{K_m + S} \right) \quad (2)$$

Monod's model has extensively been used for the estimation of biokinetic constants for bacterial growth on non-inhibitory substrates and for inhibitory substrates.

Nitrobenzene exhibits substrate inhibition and the fact that metastable conditions existed in the ASP lead to a review in the application of Monod's model to inhibitory growth models; resulting in the selection of Haldane's equation:

$$\mu = \mu_{\max} \left[\frac{S}{S + K_i + \left(S^2 / K_i \right)} \right] \quad (3)$$

where K_i is the inhibitor constant.

3 Materials and method

Pure cultures of *Pseudomonas aeruginosa* and *Pseudomonas putida*, were used in monoculture experiments. The mixed culture consisting of *Pseudomonas aeruginosa* and *Pseudomonas putida* has been used as mixed cultures in the growth responsible studies. Activated sludge biomass from a sewage treatment plant was used as the seed for plant waste studies. The seeds were then transferred into the growth medium with respective carbon sources. Specific growth rate, μ has been measured from the exponential growth phase and used in the model equations.



The wastewater from NB plant was collected as a composite sample, transported in airtight containers, and characterized. The waste after proper dilution was sterilized with a membrane filter before it was used in the growth experiments. Compositions of the NB plant waste are given in Table 1.

3.1 Analytical procedures

Nitrobenzene was estimated by gas chromatography (Varian3700). Waste water was analyzed for other parameters according to standard methods APHA, AWWA, WPCF [7]. Growth of biomass was measured by turbid metric method (OD at 610nm) which was then calibrated against dry weight. The cultures were incubated at a constant temperature of 30°C.

3.2 Experimental observations

The synthetic waste was constituted based on proximate composition and supplemented with other essential nutrients. The specific growth rate μ derived from various initial substrate concentrations has been plotted against initial constant concentrations to derive the kinetic constants using the kinetic models. Fig. 1(a)-(d) represents the reciprocal plot deriving the kinetic constants by Monod's model while Fig. 2(a)-(d) represents the typical line of best fit by Haldane's model.

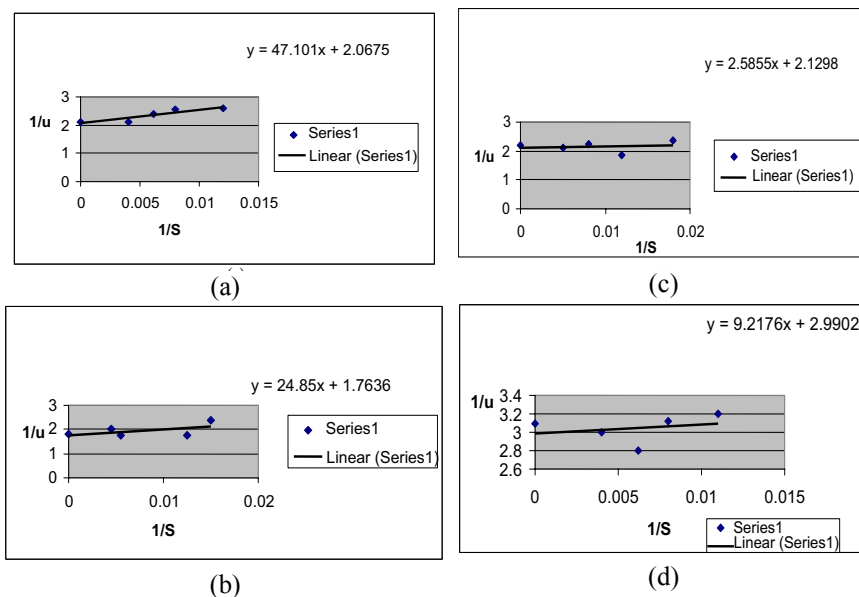


Figure 1: (a) Kinetic constants Monod model K , u_{\max} , $P. aeruginosa$. (b) kinetic constants Monod model k , u_{\max} $P. putida$, (c) kinetic constants Monod model K , u_{\max} , mixed culture synthetic waste and (d) kinetic constant Monod model K , μ_{\max} , mixed culture, plant waste.

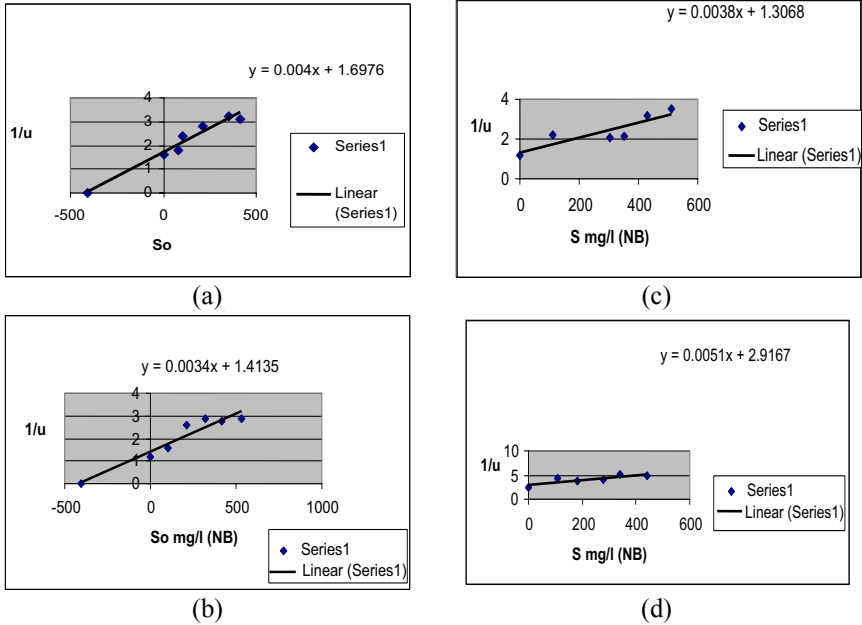


Figure 2: (a) Kinetic constant Haldane model K , μ_{\max} for *P. aeruginosa*, (b) kinetic constant Haldane model K , μ_{\max} for *P. putida*, (c) kinetic constants Haldane model K , μ_{\max} , mixed culture, synthetic waste and (d) kinetic constant Haldane Model K , μ_{\max} , mixed culture plant waste.

Table 2: Biokinetic constants for NB biodegradation of the values with Monod's and Haldane's models.

S.No	Parameter	Monod's value		Haldane's value	
		$\mu_{\max} (h^{-1})$	$K_s (mg l^{-1})$	$\mu_{\max} (h^{-1})$	$K_i (mg l^{-1})$
1	<i>Pseudomonas aeruginosa</i>	0.52	31.2	0.62	480
2	<i>Pseudomonas putida</i>	0.66	19.5	0.83	490
3	Mixed culture with Synthetic NB	0.55	12.06	0.76	440
4	Mixed culture with plant waste	0.68	103.52	0.62	380
5	Sewage plant sludge with plant waste	0.42	32.25	0.45	370

These expressions clearly indicate that μ_{\max} is a reaction rate constant achieved by a culture when it is grown under no limiting growth conditions. The value should then be the one that is obtained as the highest value. Thus μ_{\max} calculated by Haldane's model should be considered the true value.

Growth data from a continuously operated completely mixed activated sludge (CMAS) system has also been subjected to treatment using Monod's and Haldane's model equations. Experimental results obtained from batch and continuous culture experimentation has been subject to validation of mathematical models suggested by various investigators. The apprehensions and difficulties experienced in comparing the values obtained with synthetic media in nitrobenzene biodegradation and in actual waste water systems have been examined in the present study. One of the major difficulties experienced by previous investigators is the lack of techniques and tools in conducting pure culture experiments. Isolation, enrichment and monoculture experimentation of laboratory stored cultures have been possible in the author's laboratory to study the kinetics of biodegradation of nitrobenzene in absence and presence of secondary inhibitors. These studies have been found necessary to understand the maximum degradation rate (the maximum substrate utilization rate) of the substrate so that in the event of the absence of all environmental pressure how far the organisms can express themselves. The μ_{\max} for *Pseudomonas aeruginosa* of 0.62 hr^{-1} could not be achieved when the culture is mixed with other cultures and cultivated in a defined medium along with other strains of bacteria. The μ_{\max} is further reduced when the mixed cultures are grown in actual wastewater ($\mu_{\max}=0.76 \text{ hr}^{-1}$), indicating that there are several factors which influence the growth of microorganisms in a given environment. Competition for the common substrate is one of the possible reasons for a lower growth rate while noncompetitive inhibition secondary inhibitors are another. Laboratory-stored cultures are highly induced having multiple copies of plasmid genomes, which can express themselves, so that high efficiency can be achieved with such specialized seeds. Treatment of kinetic data with various model equations shows that Monod's model has limitations when used with inhibitory compounds. When the substrate induced cultures are grown at high initial substrate concentrations they tend to show lower growth rates (observations by Hill and Robinson, [8]; Gaudy *et.al*, [9]). This means that the organisms are under the inhibitory effect of the substrate. It is possible to achieve the exponential growth very fast and reproducibility of the experimentation is possible when the cultures are grown with initial substrate concentration is below the K_s . for experimentation on any kinetic studies with inhibitory constituents, it is necessary to obtain the K_s of the system, in the first instance and then with this as the critical concentration further experiments have to be conducted to obtain value for K_s . Biodegradation can considered a feasible and reasonable system to eliminate, NB, NT from the effluents (Swamy *et al* [10]).



4 Conclusions

Analysis of data on continuously operated reactor systems shows that under feed starve conditions μ_{\max} is much lower than the batch fed systems, indicating high detention time has to be given in the reactor. However, for the treatment of nitrobenzene, conventional ASPs with a HRT of 12 to 24 hrs is optimum. K_i of the continuously operated system is low: a situation that was postulated by Monod. The death rate of the organisms is very low which is ideal for detoxification of the toxic substrate. Engineering significance of kinetic constants is that, when hydraulic shock is effected beyond μ^* there is every possibility of biomass being washed out. Similarly, when organic shock is made beyond K_i value the rate of degradation of the substrate is retarded, resulting in substrate build up which also reduces the treatment efficiency and functioning of the treatment plant.

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